<u>Removal of Pphenol removal from hyper-saline wastewater using fluidized catalyst (catalytic?)</u>-bed

reactor

ABSTRACT

This paper discusses the processes involved in the successful synthesis of e-Cu/Mg/Al-chitosan was synthesized successfully-and its use in a used in a fluidized catalyst (catalytic?)- bed reactor in ord to degrade phenol from hyper-saline wastewater. The results showed that the phenol could can be completely oxidized by Cu/Mg/Al-chitosan--H₂O₂ within 7 min at acidic pH. The influence of various variables—was—investigated,__-including solution pH, salinity concentration, H₂O₂ concentration, and Cu/Mg/Al-chitosan quantity is investigated, to study their for their any effects on phenol degradation in a synthetic saline wastewater. The maximum degradation of phenol was iwas_achieved at pH 2 and 7 g Cu/Mg/Al-chitosan. During this process, it is noted that Cghloride and sulfate ions had have a synergistic effect on phenol removal, where the phenol-oxidation rate of phenol in the presence of sulfate ions is seen to be twice o times more than of that e oxidation rate-obtained in-under controlled conditions. Finally, it is found that an iIndustrial wastewater containing phenol,- maycan be could be effectively treated using a relatively low concentration of Cu/Mg/Al-chitosan, 5 g, and in a short hydraulic retention time of 7 min. Overall, the presented method demonstrated demonstrates efficiengcy and holds promise a promising as a simple and elegant method to eliminate the phenol from wastewater.

Keywords: Phenol, Degradation, Cu/Mg/Al-chitosan, Catalyst, Hyper-saline wastewater.

1. Introduction

Using Chemical processes (and their use) chemical processes in particular advances favor oxidation as an alternative method by which organic compounds may be oxidized for oxidation of organic compounds and converting converted those products into simple and minerals. These processes are highly useful necessary [1] and have recently received been the subject of significant attention. Amongst the various

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oxidation techniques known advances oxidation techniques, catalytic oxidation appears to be a promising field of study. It has been reported to beis effective for the near-near-ambient degradation of pollutants, because as it can provide holds promise of a nearly-_complete degradation [2]. Zhou et al. [3] pointed out that in recent years to there is a considerable the interest has been expressed in the development of copper-based heterogeneous catalysts, especially hydrotalcite-like compounds in recent years. Hydrotaleite like These compounds, referred to as layered double hydroxides, are a classes of layered materials and have receiveding an increasing increasing attention in recent years owing to their diverse applications, especially in catalysis. and has These may be denoted by a general formula of: CuM²AlCO₃ $(M^2 = Co^{2+}, Ni^{2+}, Cu^{2+}, Mg^{2+}, Zn^{2+}, and Fe^{2+})$. According to the literature, which has been recently reviewed by Nawrocki and Kasprzyk-Hordern [4], tThe catalysts' applicability catalysts may be affected [4] by several factors, including: leaching of metals into the liquid phase, high $cost_{2}$ of production, and the availability of catalysts in solution (*i.e.* needing to low-density supporting agents). These challenges technically, environmentally, and economically limit their most full-scale applications of catalysts. Thus Therefore, the main concern with regard to catalysts facing catalysts issurrounds the development of a more environmental-friendly catalyst with entailing a simple and low-cost production method. In this regard, ongoing-research is on has been Accordingly, research is ongoing toattempting to find novel materials with the high catalytic activities that are also may be easy and cheapeconomical and easy to produce.

Therefore, in the present study, we focus on Cu/Mg/Al (CMA) as <u>the</u> catalyst, because all elements in this compound are routinely used in waterworks <u>systems</u> and <u>are easily</u> available <u>rather thancompared to</u> Ni, Pd, and Ag that <u>may have beenwere</u> previously <u>applied preofferred</u> by other researchers [3,_4]. For example, Zhou *et al.* [3] used Cu/Ni/AlCO₃ for phenol degradation from aqueous solutions. But in this study, we <u>shall applied apply</u> Mg instead of Ni because givenas if magnesium <u>being is non-toxic</u>, and does not pollute water when used as a leaching agent in the leached into the liquid to be treated. <u>under treatment is not toxic and would not cause water pollution</u>. Moreover, for purpose of availability of catalyst in solution the chitosan compound was will be used as a supporting agent or as a the-catalyst in

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<u>solution</u>used as supporting. The e<u>C</u>hitosan has three functional groups, *i.e.* two hydroxyl groups (–OH) and one amino group (–NH₂), per glucosamine unit [5] <u>which</u> for maintaining the catalysts and <u>iIt is</u> characterized by low densities and <u>is</u> density is low for purpose of availability and covering the complete of solution. In addition, the chitosan that <u>is</u> obtained from fishery-waste <u>is much more economical base is</u> very cheaper than activated carbon, which is often used as a support agent, that usually used as supporting.

Although several works have appeared been published so far on on the use of using hydrotalcite-like compounds for the degradation of different pollutants, it is seen that anno investigation of this catalyst into theis catalyst's role as in terms of an agent to remove removal of pollutants from real saline real wastewater₂ is <u>seenvery rare</u>. <u>L</u> Based on our best literature review efforts, survey did not find, except for a few reports available on using advanced oxidation processes for the removal of phenol from saline wastewaters [6,7], any there was a marked absence of no-investigations could be found on CMA treatment of saline wastewaters containing high phenol concentrations. An The investigation on of the TiO₂-photocatalytic process for the degradation of phenol in saline solution showed a significant inhibitory effect for at 50 g/L of NaCl on the phenol degradation [6]. It was also stated that tThe time required for the effective degradation of phenol increased with as the increase in NaCl content-increased. investigated by-Maciel et al. [7] investigated, Tthe effectiveness of-the Fenton and photo-Fenton processes on phenol degradation from caused by a saline effluent was, investigated by Maciel et al. [7]. Their results demonstrated show that although both processes were effective in for phenol degradation, the high salt concentration inhibited the oxidation reaction considerably, so that, and only a-50% removal of TOC (total organic carbon) was removed achieved in the photo-Fenton process in the presence of 50 g/L NaCl, even after a reaction time of 100 min-reaction time. Moussavi et al. [8] reported an integrated system (catalytic ozonation/-biological processes) to formemove phenol removal from a saline solution. Although, they had achieved to high efficiency of removal, bulk of many of their materials and methods are prompted questionables on their from the perspectives of environmentComment [Editor5]: Please check for clarity

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friendliness and safety but many material and methods was used which some of them environmentally and safety point of view is concern.

Therefore, the present study represents the first application effort at using of fishery waste-based agents as supporting supports of for the CMA to enablefor complete removal of for removal of phenol from the hyper-saline wastewater using a fluidized bed reactor. Effects The effects of the following se basic variables like were evaluated in the tests on phenol degradation; \pm solution pH, CMA-chitosan quantity, types of salinity, and H₂O₂ concentration; were evaluated in the tests on phenol degradation. For the latter (???), the efficiency of the CMA-catalyst-H₂O₂ was investigated to study in terms of its how effectiveness it was in the to remove al of the removal of phenol from industrial hyper-saline wastewater under optimized conditions.

2. Materials and methods

2.1. Materials

The sShrimp shell waste of <u>the</u> *Philocheras lowisi* was collected directly from the Persian Gulf.<u>and</u>-The deacetylatedion was performed using a method similar to the method <u>one</u> reported by Novikov [9]. The eChitosan (is it shell waste?) was finally sieved in the size range 0.1_0.2 mm. Other chemicals and reagents used in this work-were of analytical grade and applied without further purification. Double_distilled water was used to prepare all solutions.

2.2. Preparation of the CMA-chitosan

The CMA_-chitosan was synthesized by employing metal nitrates and Na₂CO₃/NaOH. The preparation was <u>performed-readied</u> in a 250--mL flask containing metal nitrates of Cu²⁺ (0.15_-0.28 mol), Mg²⁺ (0.07_ -0.22 mol), and Al³⁺ (0.09_-0.25 mol) to <u>make-achieve a-the</u> desired Cu:Mg:Al molar ratio. Specifically, 10 g of chitosan particles <u>was-were</u> added to this solution and then_{$\overline{2}$} 250 mL-base solutions with NaOH (0.8 mol) and Na₂CO₃ (0.05 mol) were added <u>drop-drop-</u>wise into <u>the</u> flask <u>which-wasand</u> stirred vigorously with a magnetic stirrer and <u>was keptmaintained</u> for 4 h at 45-°C using a thermostated water

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bath. After that, the mixture was continuously stirred and kept at 45 °C for 4 h, <u>_It was</u> then was cooled at room temperature and filtered. The achieved solid was washed using double_-distilled water until <u>it</u> <u>became</u>_nitrate-free, <u>_. This wasand</u> then_{\overline{x}} dried at 50–°C for 7 h. Table 1 he_details of the physical and chemical characteristics of CMA_-chitosan are presented in Table 1.

2.3. Experimental procedure

The experiments were carried out in a fluidized bed reactor (FBR). As shown in Fig-<u>ure</u> 1, a glass column with <u>havingwith a</u> diameter of 20 mm and length of 250 mm was employed as the reactor. The total volume of the reactor was 78 ml. A circulation pump was installed to <u>maintain thean for maintaining the</u> up<u>ward</u> flow velocity of at least 50 m/h to fluidize-the CMA-chitosan. The hydraulic retention time (HRT) was changed by varying the flow rate of the influent and effluent pumps. The solution pH was adjusted to <u>the</u> designated values by adding 0.1 N HCl and NaOH solutions to the reservoir of raw wastewater. The reservoir was magnetically stirred and kept-maintained at the desired temperature (4-45 °C), and <u>The</u>H₂O₂ (0.02-1.02 mol/L) was added at once,-<u>and this</u>which-initiatinged the reaction. Effluent was withdrawn continuously from the top of the reactor, and 5 mL aliquots were withdrawn from their effluent as sample at designated time intervals. It was of using 0.22-um membranes to analyze the reaction mixture. The withdrawn point of the effluent was withdrawn is from a point about round-3 cm above that of the circulation pump to avoid the carry over <u>effect</u> of CMA_-chitosan.

2.4. Analysis

The pPhenol concentration in the supernatant was determined using DIONEX Ultimate 3000 highperformance liquid chromatography (HPLC). The intermediate compounds were monitored by UV—Vis spectroscopy and HPLC. In addition, the total organic carbon (TOC) was measured by a Shimadzu TOC-5000 Analyzer (Shimadzu Co., Japan). The nature of the CMA—chitosan nature was verified through Fourier transform infra-red analysis (FTIR) (Prestige, 21210045, Japan). The Chitosan and CMA— **Comment [Editor12]:** Do you mean a thermostat operated water bath?

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chitosan samples were used before the process of phenol removal. was undertaken. The X-ray diffraction (XRD) patterns were also-determined on a Rigaku D/MAX 2200 (Tokyo, Japan) instrument.

3. Results and discussion

3.1. CMA_-chitosan characterization

As shown in Figu. <u>re</u>2, <u>tThe</u> FTIR spectroscopy (Figure 2) confirms the presence of CMA in our sample and along with the interaction between the chitosan and catalyst crystal surfaces. Fig. <u>ure</u>2 depicts t<u>The</u> FTIR spectra-<u>show of</u> free chitosan and catalyst_-chitosan particles. The resemblance <u>in between theof</u> the spectral features <u>confirmed-confirms the</u> successful attachment of CMA <u>onto-to</u> the surface of the chitosan particles. The absorption bands of the <u>____OH and ____NH</u>2 stretching modes at 3393 1/cm and 1647 1/cm <u>are seen to</u> undergo discernible shifts <u>when we comparing compared the</u> chitosan <u>alone-on its</u> own with the catalyst-chitosan, <u>which</u> indicatesing a weak interaction <u>of between the</u> chitosan with and the particle surface. The <u>s</u>Complex <u>formation</u> between an amino group and CMA is most likely to take place in monodentate mode which <u>will expectedly</u> leaves more space on the surface of <u>the</u> CMA. Although <u>bB</u>oth ____NH₂ and ___OH groups of chitosan may be involved in the interactions with the CMA particle surface. The particle stabilization <u>characteristics that are derived because of from their</u> stronger binding strength with metals. These evidence<u>ss indicated show</u> that <u>CMA has modified</u> chitosan.

was has been modified by CMA.

FXRD provided further evidence for the formation of CMA was obtained by through the use of XRD (as demonstrated in Fig.-ure 3). The CMA_--chitosan in its crystalline form exhibits many sharp diffraction peaks between $2\Theta = 6-45^\circ$, while no such peaks are visible in the XRD of chitosan alone, which may occurbe because theof the trapping of chitosan is trapped by the CMA due to the trapping of CMA onto chitosan.

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3.2. Influence of Cu:Mg:Al molar ratio on phenol degradation

The Literatures reviews have search showed shows that the predominant products of catalytic phenol degradation are hydroquinone, p-benzoquinone, formic acid, acetic acid, and fumaric acid [3, 10]. Several byproducts were formed (Table 2) during the degradation of phenol. Table 2 showed shows the percentages of the by products of phenol degradation by achieved by resorting to putting the CMA-chitosan in <u>the FBR reactor along with the d</u>etermined HRT. It was found that <u>the Cu/Mg/Al</u> molar ratio influencesd the oxidation and theas also refore, the deep and catalytic activity of the catalyst. As shown in Table 2, iIncreasing Al concentrations-could significantly enhance (Table 2) -the deep oxidation of phenol into smaller molecules such as, formic acid, acetic acid, and fumaric acid owing to the presence of more surface oxygen species [11], while increased copper concentrations enhance, whereas the catalystic activity was could be enhanced by an increase in copper concentrations. InAt low Al concentrations, *i.e.* Cu/Mg/Al molar ratio of 4:2:1, aluminum did does would not play any important role and therefore, Cu/Mg/Al had-becoogemes a the bimetallic of Cu/Mg, In the Cu/Mg bimetal, as stated in our a previous study [12], the rapid phenol conversion of phenol takes place, probablymay be explained by due to mechanisms such as_x the formation of metal_-hydride complexes with the copper and the dissociation of molecular hydrogen or other hydrogen sources on the surface of copper serving as a direct reductant for phenol. Deeply degradation of phenol was of our goal; hencetherefore, we selected the ratio 2:2.6:2 ratio for further experimentation.

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3.3. Influence of pH and synergy of CMA, H_2O_2 , and chitosan

Given the effect of Due to affecting the charge distribution on the catalysts' surfaces [13], as well as and the pathways and kinetics of the catalytic reactions, the solution pH wais seen to-playing an plays an important role in the overall performance of the CMA-chitosan-H₂O₂. Therefore, a series of experiments were carried-outconducted to evaluate the catalytic ability of the prepared CMA-chitosan particles at different pHs in phenol degradation in saline wastewater. Fig.-ure 4 presents the time-course of phenol degradation at several solution pHs ranging from 2 to 10. As seen in Fig.-ure 4the figure, phenol the degrades ation of phenol was-the highest at a pH of 2_7 with here-98.7% accuracy of phenol was removed

after a 7-min retention time, and the lowest- at alkaline pH. In order tTo better illustrate the effect of pH, the kinetics of phenol degradation was assessed for CMA-chitosan-H₂O₂. The resulting information data is has been in summarized in Table 3, which indicates shows that the oxidation of phenol under the selected conditions was iwas of pseudo- first -order. As seen in Table 3, tThe reaction rate constant was decreased progressively decreased from 0.028 to 0.012 l/min (corresponding to maximum degradation of phenol from 100 to 56.5%) when the pH was increased from 2 to 10. Attainment of tThe maximum phenol degradation at acidic pH $\frac{\text{may}}{\text{may}}$ be_- $\frac{\text{attributed} \text{due}}{\text{may}}$ to the complex interaction $\frac{\text{of}-\text{between}}{\text{between}}$ H₂O₂ and phenol molecules with the catalyst surface., which This will be discussed further later in the next section. The improvement of in the phenol degradation rate with decreasing pH might may be due attributed due to the acceleration of H_2O_2 mass transfer and H_2O_2 decomposition rates with pH [14]; this in turn is seen to have leads to the formation of highly reactive radicals, mainly 'OH [15]. The A decreasing decrease in the pH₃ the higher would beis the leads to higher OH produced OH and thus therefore, to he higher is the degradation rate. The solution pH had a negative effect on phenol degradation at values 8 and higher $\frac{1}{4}$ *i.e.* the degradation rate of phenol was-reduced under strong alkaline conditions (Table 3). This finding canmaycan be better interpreted by considering the taking into consideration both the speciation of phenol (pK_a) and the surface charge of the CMA-chitosan, with respect to solution pH and CMA-chitosan pH_{zpc} . Since the pH_{zpc} of CMA-chitosan is 7.8, a negative charge is developed on its surface at these pHs (8 and 10). Phenol, on the other hand, is mostly dissociated into phenolate anions ($pK_a = 9.9$) at this basic ity conditions- [8]. Therefore, the affinity of CMA-chitosan for phenolate under strong alkaline conditions is restricted, and the surface catalyst reactions would are most likely to be probably be inhibited, leading to the reduction in phenol degradation rate. Moreover, the low phenol removal in alkaline solution is can be <u>due-attributed</u> to <u>the</u> instability of H_2O_2 [3]. In alkaline pH, H_2O_2 immediately decomposes to-produce H_2O and O_2 , accordingly with <u>a loss of in</u> oxidizing ability-loss.

The degradation rate increased increases with decreasing pH_{τ} due toand this may be justified. This can be justified by the fact that the decomposition of H_2O_2 which increases with decreasing pH_{τ} resulting in the formation of highly reactive •OH [3,16] with high oxidation potential of 1.8 V.

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For a better illustration of the catalytic role of the prepared CMA in CMA-chitosan-H₂O₂, an experiment was carried out in which a-CMA-chitosan saturated with phenol was used instead of a fresh one (Figure 5). The results are shown in Fig. ure 5. As can be observed in Fig. 5here, the percentage of phenol removedal percentages was about the same with both prepared and fresh did not deviate significantly from those obtained with the fresh CMA-chitosan-H₂O₂ under at the same imilar experimental conditions-This suggestings athe catalytic role is the prevalent reaction in the to CMA-chitosan-H₂O₂ rather than that <u>of</u> adsorption—oxidation—one, <u>Itas also</u> and <u>indicates</u> a synergistic <u>role</u>, <u>effect of CMA with H_2O_2 -in</u> degrading the phenol. According to Fig. ure 5, aAround 25% of the phenol was adsorbed onto the CMAchitosan during a short retention time of 7 min_(Figure 5); thereafter, the adsorption percentage remained almost unchanged up to an-a contact time of 105 -min-contact time. This reveals the low capacity and rapid adsorption properties of phenol on CMA-chitosan particles and thus the saturation of the catalyst. The degradation of phenol in CMA-chitosan-H₂O₂ commenced atwas started by 93.9% for-during a contact time of a 7 -min-contact time and increased to 100% after a-a period of 70 -min-contact time. It is worth noting that the pH of the solution in the CMA-chitosan-H₂O₂ descended from an initial value of 6.5 to around 3.5 after a reaction time of 7-min-reaction time. This is could be may be due to the formation of some acidic intermediates such as₃ oxalic, acetic, and formic acids during the oxidation reaction [17]. Our studies are in agreement with This is was in accordance with Moussavi et al. [8] who, who have shown found a higher degradation of the phenol in aqueous solutions employingusing ozonation- and with MgO. To quantify the contribution of the prepared CMA-chitosan used in the CMA-chitosan- H_2O_2 in degrading phenol in saline wastewater was quantified by calculating, its he synergistic influence-achieved from adding the CMA chitosan toon the reactor in the presence of H_2O_2 , was calculated It is quantified as follows: F

<u>sSynergistic</u> influence = <u>pP</u>henol removed <u>infrom</u> CMA-chitosan-H₂O₂—–(<u>sumQuantum</u> of phenol removed by H₂O₂ and adsorbed onto the CMA-chitosan)]_, and the results are also<u>were plotted in (Fig-ure</u> 5).

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As <u>is</u> evident <u>in this figurehere</u>, <u>combining the prepared CMA-chitosan with H₂O₂ had</u> a significant synergistic influence (38% for<u>in</u> a 7___min contact time and 64% for<u>in</u> an 140___min reaction time) was attained from combining the prepared CMA chitosan with H₂O₂-ion phenol removal. This finding <u>confirms</u> verifies the strong ability of the prepared CMA to catalyze the H₂O₂ reaction for<u>in</u> the degrading attom of phenol in saline wastewater. The high <u>percentage in</u> removal <u>percentage</u> of phenol <u>by</u> in the CMA-chitosan-H₂O₂ <u>can may be accounted forshows by</u> the capability of CMA to decompose H₂O₂ and thereby_x enhance •OH generation [18]. Some researchers [*e.g.* 8, 19] have reported the synergistic effect <u>infor</u> phenol removal.

3.4. Influence of CMA-chitosan and H₂O₂ concentration

The phenol removal e<u>E</u>xperiments were <u>conducted</u>performed at CMA chitosan concentration (1-7 g) to determine the <u>ideal</u>contributions of <u>the</u>_catalyst-chitosan dosage <u>for</u> in phenol degradation. Results <u>The</u> <u>rResults</u> showed (Fig. 6a) that <u>the</u> phenol degrad<u>edation increased</u> rapidly <u>at higher with increasing</u> CMA-chitosan concentrations. The <u>hH</u>igher the amounts of CMA-chitosan, the provide higher <u>wais the</u> amounts of <u>the</u> catalyst <u>formed/required?</u>. Increase <u>An</u> increase in the <u>eatalyst catalyst _dosage madeled to</u> more surfaces <u>becoming</u>_available for the reaction with H₂O₂ [10], subsequently resulting in <u>an</u>-increased phenol oxidation. Phenol removal is complete at <u>W</u>, wherein, when the <u>a</u> concentration of CMA_-chitosan <u>concentration of reached to between 5 and 7 g., the phenol could be removed completely. These <u>Results demonstrateshowd</u> that <u>at in</u> higher dosages of catalyst related to the, greater availability number_of catalyst surface active sites to react with the morelarger quantity of H₂O₂; these increases therefore enable enhanced generation of reactive radicals, mainly •OH, resulting in improvement of the in phenol removal efficiency.</u>

In another attempt, the effects of H_2O_2 concentration on phenol degradation were was tested (and the results are were shown in Fig.-ure 6b). When the concentration of H_2O_2 concentration increased from 0.02 to 0.06 mol/L, the phenol degradation percentages also increased correspondingly from 28% to 98.2%. The increasing in H_2O_2 -concentration in the solution leads to an increase in the formation of •OH [20].

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However, when given<u>at</u> higher H_2O_2 dosages, the phenol removal was not further enhanced did not-get enhanced further. On To the contrary, it dropped downbut dropped down. This is wais because, at high H_2O_2 concentration, both the substrate and H_2O_2 there is was a competed ition of for •OH-consumption between the substrate and the H_2O_2 . As reported by Mingee-Long *et al.* [21] state that, H_2O_2 atim high concentration could acts as a scavenger of the highly potent •OH, and recombine with •OH would recombine to form H_2O and O_2 . Therefore, it would be necessary to select an optimal H_2O_2 concentration for catalytic reactions.

3.5. Influence of salt type

The influence of sSome inorganic salts including NaCl, KCl, KNO₃, and Na₂SO₄ oninfluence phenol conversion is summarized in (Table 4). It is clear that the pre_sence of saltsand affects the performance of CMA-chitosan-H₂O₂ in different waysly with respect to phenol oxidation. The presence of NaNO₃ markedly decreased the conversion rate of phenol markedly, whereas the presence of chlorides and sulfate remarkably-increased the oxidation rate remarkably. This means that-is meaning that the NO₃ actsed as a radical scavengers to suppress the degradation of phenol, in which their its inhibitory effects being awas attributed to a decrease in the rate of generation of hydroxyl radicals (•OH). As it can be seen from Table 4, <u>chloride ions increase the phenol</u> oxidation rate and <u>decrease</u> HRT in the presence of chloride ions was increase and decrease, respectively, in comparison contrast to control, i.e. noabsence of any salt, where the opposite is trues (control condition), (???) De Laat et al. [16] previously-stated that the chloride ions was wereact as radical scavengers resulting in and caused the formation of radicals such as $\circ Cl_2^-$ that weare less reactive than the OH. Therefore, this novel phenomenon requires d more than just a cursory glanceattentionit deserves to pay much attention to the novel phenomenon. Vione *et al.* [22] pointed out that phenol chlorination might occur in the presence of dissolved Fe³⁺, hydrogen peroxide, and chloride. Furthermore, halogenated quinones also the decomposeition of H2O2 could be promoted by halogenated quinones, resulting in the formation of •OH [23]. GC/MS confirm aA detectable amount of chlorinated quinones or other intermediate products were confirmed in our experiments by GC/MS.

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Surprisingly, we found <u>that a significant impact of sulfate ions impacted significantly on the phenol</u> oxidation rate <u>significantly</u>. The phenol oxidation rate in <u>the presence of sulfate ions was seen to</u> <u>beconverting twice of two times more that a the oxidation rate obtained in under</u> controlled conditions. This probably means implying of a possible role for <u>ies</u> that some radicals containing sulfur <u>could played</u> an important role in promoting the peroxide oxidation of phenol when <u>the sulfate ion wais</u> added. The <u>eEfforts was were conducted made</u> to detect <u>the persulfate radical</u> ($S_2O_8^{2-}$)₇ and we successfully found <u>obtained</u> 4.8 mg/L of <u>persulfate radicalit</u>. The sulfate ion under unclear mechanisms converted to persulfate radical ($2SO_4^{2-} \rightarrow S_2O_8^{2-}$). The persulfate radical is <u>known as a very strong oxidizer and is used</u> for <u>purposes of for</u> decontamination <u>purpose</u>, <u>as given its oxidation potential</u>, at 2.1 V, <u>which is only</u> slightly <u>weaker lesser</u> than <u>that of ozone</u> (O_3^{2-} , 2.2 V) but stronger than both_r hydrogen peroxide (H₂O₂, 1.8 V) and permanganate (MnO₄, 1.7 V).

3.6. Durability and leaching the catalyst metals

An important characteristic of a catalyst, from a practical point of view, is its deactivation or durability potential. To evaluate the durability of the CMA-chitosan in the oxidation of phenol, the we used 5 g of CMA-chitosan was used in FBR with HRT of 7 min. The phenol oxidation efficiency was and determined it after every each 7 min, and results are have been plotted in (Fig.-ure 7). As indicated, the CMA-chitosan preserved its catalytic properties even after 5 h and after that until 7th h_x it was noticed that theits efficiacy energy was still good acceptable (>70%): probably This can may be described dueeither by to the predominance of catalytic reactions rather than adsorption—oxidation reactions in the CMA-chitosan, or to as the in_-situ regeneration of chitosan [24]. A closer look at Fig.-ure 7 shows reveals that the procedure of phenol removal accelerated has improved been slightly improved after 1 h, probably due to the modifications of to chitosan [24]. but decreased. However, significantly after the 7 h operation, the phenol removal processwas significantly decreased. This probably due may was attributed to the be due to

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leaching of the catalytic active species or the poisoning of the active sites or even the fouling of the catalyst surface by intermediate reaction products.

Based on t<u>T</u>he results of this study; show that di-valent magnesium and copper ean-could be released into the solution during the degradation of phenol by the CMA-chitosan-H₂O₂ system. Therefore, we made an effort was only made to identify the amounts of Cu²⁺ and Mg²⁺ ions during the experiments; as because the Al³⁺ ion was provedwas undetectable in the effluents. The results (are have been shown in Fig.-ure 7). The relatively high levels of magnesium ion accumulation (more than the MCL of Mg²⁺ ion in drinking water) was occurred, and this results resulted in the formation of in hard water. As showndepieted in Fig.-ure 7, the copper ion-leached during the experiment was continually lower than the maximum contaminantion level of Cu²⁺ ion in drinking water. Lowering the leaching amount of metal ions led to a decreased ing-the phenol removal procedure bycapability of CMA-chitosan-H₂O₂.

3.7. Industrial wastewater treatment

The <u>next experiment was to study the</u> efficacy of CMA-chitosan-H₂O₂ in the removal of phenol from industrial wastewater was <u>subsequently</u> evaluated. Ffor this which, a we obtained bulk wastewater sample was obtained from a petrochemical plant. The characteristics of the wastewater are presented in Table 5. The experiment was conducted at pH level as per <u>the</u> original wastewater with CMA-chitosan concentrations of 5 g, and HRT of 10 min. The main parameters of the wastewater sample before and after treatment with CMA-chitosan-H₂O₂ are archave been presented in Table 5. As seen in The table 5. shows that a very low quantity of CMA-chitosan at very short HRT (10 min) could completely removed phenol and improved some of <u>the</u> other characteristics of the treated wastewater. A point to be noted that can be derived from Table 5 is that <u>of the</u> TOC reduction during treatment. The CMA-chitosan-H₂O₂ not only efficiently degraded the phenol <u>efficiently</u>, but and also attained a high degree of TOC reduction leading to and thus₃ mineralization of the degradation intermediates. The oxidizing radical species attacked

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to the aromatic rings, and leading to its the degradention. Upon degradation of the of the phenol molecules, suppon which, tThe degradation intermediates are then weare-subjected to oxidized ation by the oxidizing radical agents including the hydrogen radical and other radical species, causing the a subsequent decreasinge in the TOC content further. These findings confirm the capability of the CMA-chitosan-H₂O₂ for the treatment of phenol--wastewater.

4. Conclusions

The CMA-chitosan-H₂O₂ process has an excellent <u>runperformance in the process of the</u> degraded_ation of-phenol frominfrom hyper-saline wastewater_efficiently, and it whas found to have a higher catalytic capability in the presence of chloride and sulfate ions. The oOptimal values of operational parameters including CMA-chitosan quantity, H₂O₂ concentration, and solution pH werehave also been_obtained. Furthermore, CMA-chitosan-H₂O₂ reduced the high concentration of phenol in industrial wastewater to levels around the designated standards for effluent discharge. It-mayis thus proves be concluded that the CMA-chitosan-H₂O₂ process <u>iwais</u> a promising and economically viable <u>processtechnology</u> for the treatment of saline wastewater containing phenol.

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Fig. 1. Schematic of experimental apparatus used for catalytic degradation of phenol.

Fig. 2. FTIR spectra of chitosan (a) and CMA--chitosan (b). Fig. 3. XRD patterns of chitosan (a) and CMA-chitosan (b). Comment [Editor34]: Please add the location

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Fig. 4. The influence of pH on phenol degradation from saline wastewater in the CMA-chitosan- H_2O_2 (phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Fig. 5. Phenol removal by-the H₂O₂, CMA-chitosan-H₂O₂, and adsorption onto CMA-chitosan particles (phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Fig. 6. (a) Influence of CMA-chitosan amount and (b) H_2O_2 amount on phenol degradation (phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Fig. 7. The durability of the CMA-chitosan in oxidation of phenol by CMA-chitosan- H_2O_2 and leaching of metal catalysts (phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Table 1. The physical and chemical characteristics of CMA-chitosan.

Table 2. Reaction rate constants of phenol degradation in the CMA-chitosan- H_2O_2 at various pHs (phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Table 3. Influence of Cu/Mg/Al molar ratio on phenol conversion and phenol oxidation intermediates (pH: 6, phenol concentration: 1000 mg/L, NaCl concentration: 55 g/L).

Table 4. Influence of inorganic salts on phenol conversion by CMA-chitosan- H_2O_2 (pH: 6.1, phenolconcentration: 1000 mg/L, NaCl concentration: 55 g/L).

Table 5. The quality of petrochemical wastewater before and after treatment with the CMA-chitosan- H_2O_2 .